

# Association between weeds and plant growthpromoting rhizobacteria in the phytoremediation of lead-contaminated soil



Asociación entre malezas y rizobacterias promotoras del crecimiento vegetal en la fitorremediación de suelo contaminado con plomo

https://doi.org/10.15446/rfnam.v77n2.108353

Sergio Muro-Del Valle<sup>1</sup>, Alejandro Mago-Córdova<sup>1</sup>, Carmen Carreño-Farfán<sup>1</sup>, Marilín Sánchez-Purihuamán<sup>1\*</sup>, Junior Caro-Castro<sup>2</sup> and Martin Carbajal-Gamarra<sup>3</sup>

#### ABSTRACT

## Keywords:

Bacterial consortia Echinochloa colona (L.) Link Lead tolerance Phytotoxicity Synergy Lead is a persistent heavy metal in the soil that can accumulate in edible plants, so non-polluting strategies are required for its removal. In this study, the efficiency of weeds with associated rhizobacteria in phytoremediation of soil contaminated with lead (800 ppm) was investigated. Weeds with lead tolerance were selected, as well as rhizobacteria that promote plant growth in vitro. Several bacterial consortia (BC) were applied on three weed species, and the weight of the aboveground biomass of the weeds, the phytotoxicity of the soil after phytoremediation, as well as the parameters of the phytoremediation of lead in the soil with lower phytotoxicity, were evaluated. As a result, 20% of the weeds analyzed were tolerant to lead with indices of 0.80 (Echinochloa colona (L.) Link), 0.76 (Cyperus corymbosus Rottb.), and 0.72 (Sorghum halepense). BC solubilized phosphates, produced indole acetic acid, and increased the fresh biomass of plants (4.14-14.32%). Furthermore, the lowest level of phytotoxicity in the soil was detected in the treatment of E. colona (L.) Link with Pseudomonas spp. and Acinetobacter spp. (BC1), as well as a bioaccumulation factor of 0.1650 in the foliage, 1.0250 in the roots, and a translocation factor of 0.1611. Finally, 78.83% lead removal was determined in E. colona (L.) Link with rhizobacteria, compared to the 57.58% obtained with E. colona (L.) Link without rhizobacteria. The efficiency of the association of weeds and plant growth-promoting rhizobacteria in the phytoremediation of soils contaminated with lead was demonstrated.

#### RESUMEN

Palabras clave:
Consorcios bacterianos
Echinochloa colona (L.) Link
Tolerancia al plomo
Fitotoxicidad
Sinergia

El plomo es un metal pesado persistente en el suelo que puede acumularse en las plantas comestibles, por lo que se requieren estrategias no contaminantes para su remoción. En este estudio se investigó la eficiencia de malezas con rizobacterias asociadas en la fitorremediación de suelo contaminado con plomo (800 ppm). Se seleccionaron malezas con tolerancia al plomo, así como rizobacterias que promueven el crecimiento vegetal in vitro. Se aplicaron diversos consorcios bacterianos (CB) sobre tres especies de malezas, y se evaluó el peso de la biomasa aérea de las malezas, la fitotoxicidad del suelo después de la fitorremediación, así como los parámetros de la fitorremediación del plomo en el suelo con menor fitotoxicidad. Se encontró que el 20% de las malezas analizadas fueron tolerantes al plomo con índices de 0,80 (Echinochloa colona (L.) Link), 0,76 (Cyperus corymbosus Rottb) y 0,72 (Sorghum halepense). Los CB solubilizaron fosfatos, produjeron ácido indol acético y aumentaron la biomasa fresca de las plantas (4,14-14,32%). Además, el menor nivel de fitotoxicidad en el suelo se detectó en el tratamiento de E. colona (L.) Link con Pseudomonas spp. y Acinetobacter spp. (CB1), así como un factor de bioacumulación de 0,1650 en el follaje, 1,0250 en las raíces y un factor de translocación de 0,1611. Finalmente, se determinó 78,83% de remoción de plomo en E. colona (L.) Link con rizobacterias, a comparación del 57,58% obtenido con E. colona (L.) Link sin rizobacterias. Se demostró la eficiencia de la asociación de malezas y rizobacterias promotoras del crecimiento vegetal en la fitorremediación de suelos contaminados con plomo.

<sup>1</sup>Laboratorio de Investigación Biotecnología Microbiana, Facultad de Ciencias Biológicas, Universidad Nacional Pedro Ruiz Gallo, Perú.

sergio.muro@unprg.edu.pe ២, amago@unprg.edu.pe ២, ccarreno@unprg.edu.pe ២, msanchezpu@unprg.edu.pe

<sup>2</sup>Laboratorio de Ecología Microbiana, Facultad de Ciencias Biológicas, Universidad Nacional Mayor de San Marcos, Perú. carocastrojunior@gmail.com <sup>3</sup>Energy Engineering, University of Brasilia, Brazil. fcarbajal@unb.br <sup>1</sup>

\*Corresponding author



ore than 90% of environmental pollutants, such as heavy metals, are retained in soil particles. Lead contamination is significant because this heavy metal has a long residence time, persisting for 1000–3000 years (Rodríguez et al. 2016). Environmental exposure to lead (1,300–32,260 mg kg<sup>-1</sup> in soil) affects public health worldwide (Covarrubias and Peña 2017) and is the cause of 143,000 deaths per year (Rodríguez et al. 2016). Lead (0.13-446 ppm) has been quantified in irrigation water contaminated with industrial residues or remains of fertilizers, in vegetables eaten fresh, strawberries, potatoes, cassava, sugar cane juice, and industrial fruit juices and vegetables (Salas-Marcial et al. 2019), a condition that constitutes a potential problem in food safety for humans and animals.

Lead enters the body through the respiratory, digestive, and dermal routes (Covarrubias and Peña 2017). Lead is not essential for plants and does not have any biological function, but it is transported to plant tissues. Once inside, the heavy metal increases the production of reactive oxygen species (ROS) that affect lipid cell membranes and reduce photosynthetic activity, with symptoms such as chlorosis, necrosis, decreased root and foliage growth, as well as changing root branching patterns (Nakbanpote et al. 2016).

Resistant plants present mechanisms such as evasion and tolerance that allow them to survive with high concentrations of heavy metals. In evasion, root cells limit and restrict the uptake and movement of heavy metals in plant tissues through the mechanisms of root uptake, metal precipitation, and exclusion. Exclusion barriers between roots and foliage minimize the accessibility of metals present in the soil to the roots (Sabreena et al. 2022). When the mechanisms are not adequate to reduce toxicity in plants, ROS are formed in the cytoplasm, which causes oxidative stress. In this context, plants activate enzymatic and non-enzymatic antioxidant defense mechanisms to overcome the negative effect of ROS (Sabreena et al. 2022).

Phytoremediation is defined as the use of plants and associated microorganisms to reduce contaminants in water, soil, and air. Also, It does not contaminate the environment and can be used to extract or immobilize metals, metalloids, and radionucleotides, as well as organic xenobiotics (Rigoletto et al. 2020; Sabreena et al. 2022). However, in some cases, the remaining concentrations exceed the maximum permissible limits established by the regulations to consider it as bioremediated soil (MINAM 2017). Phytoremediation can be improved with strategies that favor and accelerate the process, such as the synergy between plants and plant growth-promoting rhizobacteria (PGPR). These plant-associated microorganisms modify the bioavailability of lead through the production of siderophores, organic acids, exopolysaccharides, and solubilization of precipitated phosphates (Shah and Daverey 2020).

Weeds are fast-growing and can be used for phytoremediation of soils contaminated with lead (Contreras-Pinto et al. 2016). However, before its use, it is necessary to investigate its capacity to reduce the contaminant alone, and in combination with rhizosphere microorganisms, with the perspective of overcoming the limitations that affect the process. For this reason, this study aimed to determine the potential of weeds and associated rhizobacteria for soil phytoremediation of a solid waste dump contaminated with lead.

# MATERIAL AND METHODS Sampling location

The municipal solid waste dump, located in Lambayeque, Peru, covers an area of 2.95 km<sup>2</sup> between the parallels 06° 55' 04" South latitude and 79° 44' 07" West longitude. An identification sampling (IS) was carried out, aimed at determining if the soil was contaminated in an area of 1 ha (0-10 cm depth). The sampling pattern was of random distribution with regular grids, for which nine grids or cells (333.5 m<sup>2</sup>) were delimited with parallel and perpendicular lines (MINAM 2017). In 1 m<sup>2</sup> squares delimited in the central part of each cell, the 2 cm of surface soil was removed. In each square, 24 kg of soil (10 cm deep) were collected and deposited in polyethylene bags.

The soil from the nine sacks was mixed on a blanket and homogenized using the conning and quartering method (Contreras and Carreño 2018), and two representative samples (1 and 2 kg) were collected. In the 2 kg sample, the physical and chemical analysis was performed, while in the 1 kg sample, the count of total microorganisms was performed using the most probable number (MPN) technique in nutrient broth and the MPN of lead-tolerant microorganisms in nutrient broth with 50 mg of Pb L<sup>-1</sup>. The

rest of the soil was kept in the greenhouse of the Pedro Ruiz Gallo National University for the experimental phase of this research. The soil had a sandy texture, non-saline condition (EC=0.41 dS m<sup>-1</sup>), organic matter (0.02%), nitrogen (0.006%), phosphorus (3.7 ppm), potassium (131 ppm), lead (11.91 mg kg<sup>-1</sup>), cadmium (1.12 mg kg<sup>-1</sup>) and chromium (16.94 mg kg<sup>-1</sup>).

#### Selection of lead-tolerant weeds

In six agricultural fields, six specimens of 15 weeds with similar height and at the beginning of flowering were selected. With a shovel, the plants with their roots and rhizospheric soil were extracted and then transplanted into polypropylene pots (4 kg capacity) with 3 kg of soil from the municipal waste dump, mixed with 0.3 kg of rice husks.

Two specimens of Echinochloa colona (L.) Link, Cyperus corymbosus Rottb. and Sorghum halepense were planted in the pots (three pots per weed) and irrigated with drinking water (stored for 24 h) twice a week. Weeds were classified as adapted to transplanting if their vigor and green color were maintained and they developed leaves or shoots the following 10 days. In the adapted weeds, the tolerance to lead was investigated, for which four of the six plants were irrigated with a solution of lead nitrate Pb (NO<sub>2</sub>)<sub>2</sub> at a concentration of 10 ppm of Pb once for the first, second, and third weeks. Since the fourth week, irrigation was carried out every 3 days with a Pb (NO<sub>2</sub>)<sub>2</sub> solution, whose concentration increased geometrically for five weeks: 50, 100, 200, 400, and 800 ppm of Pb. The two remaining specimens of each weed were irrigated with lead-free water and constituted the reference controls of the normal phenotype.

Lead tolerance was qualified by the survival of 100% of the plants, and the physical appearance similar to the controls, in terms of height, vigor, and color of the leaves. After 60 days of the first irrigation with the Pb  $(NO_3)_2$  solution, the biomass of the plants was weighed and the stress tolerance index was calculated according to Chaturvedi et al. (2020) method. Weeds tolerant to Pb (800 ppm) were those that presented tolerance indices of 0.5-0.8, while highly tolerant weeds presented indices greater than 0.8 according to Frachia et al. (2022).

#### Plant growth promotion by rhizospheric bacteria

The rhizospheric soil (10 g) of lead-tolerant weeds was

inoculated into 100 mL of nutrient broth and incubated at 30 °C for 5 days (Gupta et al. 2018). The bacteria were isolated using nutrient agar supplemented with 50 mg L<sup>-1</sup> of Pb, incubated at 30 °C for 5 days. The developed colonies were grouped according to their morphological characteristics (morphotypes) and reaction to Gram staining. A representative of each morphotype was selected, and those that grew (turbidity) in nutrient broth with 100, 200, 400, and 800 mg L<sup>-1</sup> of Pb were cultured in nutrient agar with and without lead, and incubated at 30 °C for 48 h, constituting the pure cultures of lead-tolerant bacteria (Manzoor et al. 2019).

The characteristics that show the *in vitro* plant growth promotion was investigated in lead-tolerant bacteria. The phosphate-solubilizing activity of bacteria was preliminarily evaluated on Pikovskaya agar by observing the appearance of a translucent halo around the colonies after 96 h of incubation, while the quantification of phosphate solubilization was evaluated in Pikovskaya broth. On the other hand, indole acetic acid (IAA) production was evaluated in Tryptic Soy Broth (TSB) plus the Salkowski reagent, according to Liu et al. (2022).

The soluble phosphorus concentration was calculated with equation 1 using a wavelength of 690 nm, and the IAA concentration with equation 2 using a wavelength of 530 nm.

$$X_1 = \frac{(Y_1 - 0.0002)}{0.07} \tag{1}$$

$$X_2 = \frac{(Y_2 - 0.076)}{0.004} \tag{2}$$

Where  $X_1$  is the soluble phosphorus concentration (ppm),  $Y_1$  is the corrected absorbance of bacterial sample 'n' minus the absorbance of the culture medium control (0.119);  $X_2$  is the concentration of IAA (ppm),  $Y_2$  is the corrected absorbance of bacterial sample 'n' minus the absorbance of the culture medium control (0.080).

### Influence of rhizospheric bacterial consortia on the aboveground biomass of weeds and phytotoxicity of the phytoremediated soil

The soil (50 kg) previously collected in the dump was sieved (<5 mm), mixed with 5 kg of rice husk, autoclaved at 121 °C for 20 min (Chaturvedi et al. 2020) and conditioned in 18 pots, at a rate of 3 kg per pot. Next, the soil of nine

pots was artificially contaminated with a  $Pb(NO_3)_2$  solution until reaching 0.127% w/w, equivalent to 800 mg kg<sup>-1</sup> of Pb, considered the minimum concentration for commercial/ industrial/extractive soil according to the Environmental Quality Standards (EQS) established by MINAM (2017).

After one month, the three lead-tolerant plant species were transplanted in triplicate into pots with lead-contaminated and non-lead-contaminated soil (two plants per repetition). Before transplanting, in the treatments where the bacterial consortia (BC) were inoculated, the roots of plants with a similar phenotype were immersed in a suspension of the corresponding bacterial consortium for 30 min.

The inoculum of the bacteria was obtained by the largescale sowing method with two volumes: mother and definitive. The bacteria were grown independently in nutrient broth with 800 ppm of Pb. The mother culture was obtained after inoculating 1.8 mL of cultured broth with each bacterium in 16.2 mL of nutrient broth with 800 ppm of Pb and incubated at 30 °C for 24 h. The definitive culture was obtained after inoculating the mother culture (18 mL) of each bacterium in 162 mL of nutrient broth with 800 ppm of Pb and incubated at 30 °C for 24 h. In this way, 180 mL of the definitive culture of each bacterium were obtained, while for the consortia, the definitive cultures of six bacteria per weed were mixed with a total of 1,080 L (Acosta and Bustamante 2020). Finally, at 90 days after transplanting, the fresh aerial biomass of each plant was weighed.

### Phytotoxicity of the phytoremediated soil

The phytotoxicity of the phytoremediated soil was determined in seeds of *Raphanus sativus* L. "rabanito" var. Crimson Giant. In Petri dishes, 10 g of soil from each repetition of the six investigated treatments were deposited in triplicate, moistened with sterilized distilled water, and 25 seeds (in 18 Petri dishes) were conditioned in each plate. Three Petri dishes with uncontaminated soil were included as control. After 120 h at room temperature (20-22 °C), the germinated seeds were selected and quantified, the length of each radicle was measured, and the relative germination percentages (RGP), relative radicle growth (RRG) and the germination index (GI) were calculated according to Lakhal et al. (2017).

The phytotoxicity level was determined by relating the GI to the presence of toxic substances: GI  $\geq$  80%, a

very low number of toxic substances with a low level of toxicity, 80%>GI>50%, a moderate number of toxic substances with a moderate level of toxicity, and GI  $\leq$  50%, an abundant number of toxic substances with a severe level of phytotoxicity. The soil and weeds, both with and without bacterial consortia, which exhibited the lowest levels of toxicity in radish seeds, were collected to quantify lead and determine phytoremediation parameters.

# Parameters of lead phytoremediation in soil with the lower phytotoxicity

The roots and foliage of the selected weeds were washed, and together with the soil that accompanied them, were dehydrated in an oven at 75 °C for 96 h (Lu et al. 2021). Both were conditioned and sent to the Soil, Plant, Water, and Fertilizer Analysis Laboratory of the National Agrarian University - La Molina to quantify lead by atomic absorption spectrophotometry (Perkin Elmer, Model PIN 500). The quantified lead values in the soil, roots and foliage were used to calculate the bioaccumulation factor (BAC), the translocation factor (TF) and the percentage of lead removal from the soil (Chaturvedi et al. 2020).

#### Identification of Phytoremediating Bacteria

The identification of the genus of bacteria in the consortium was carried out by several biochemical like catalase, oxidase, motility, production of acidity from glucose on triple sugar iron (TSI) agar, reduction of nitrates to nitrites, hydrolysis of urea and starch, indole production, citrate utilization, lysine decarboxylation on lysine iron agar (LIA) and gelatin and esculin hydrolysis (Silva et al. 1999; Salazar and Nieves 2005; Álvarez-López et al. 2014; Dipak and Sinha 2017).

### Statistical analysis

The data analyzed were represented with the mean  $\pm$  standard deviation (SD) of three repetitions. Statistical analysis of fresh biomass was performed using variance analysis (ANOVA), followed by Tukey's test (*P*<0.05). For all the analyses, the statistical program SPSS V. 22.0 was used.

### RESULTS AND DISCUSSION Selection of lead-tolerant weeds

A 20% of the selected weeds were considered tolerant to lead (800 ppm), with tolerance indices of 0.80 (*Echinochloa colona* (L.) Link), 0.76 (*Cyperus corymbosus* Rottb.) and

0.72 (*Sorghum halepense*). The tolerance of weeds to lead (800 ppm) coincides with the reports of Amadi and Gbosidom (2022) with *E. colona* (L.) Link, and Sabreena et al. (2022) with *Sorghum* spp. Also, lead decreased the fresh biomass of weeds (Figure 1), a negative effect previously reported by Chaturvedi et al. (2020). The researchers demonstrated that lead (50-100 ppm) decreased plant height (15.38%) and biomass (13.7%), as well as chlorophyll, carotenoids, and nitrogen. On the other hand, Frachia et al. (2022) reported that lead (100-500 ppm) increased the weight of foliage and roots

of *Inga uruguensis*, and based on the fresh biomass of the plants, a tolerance index (TI) of 1.21 was reached.

The morphological and physiological changes of plants in lead-contaminated soils vary according to the concentration of the heavy metal and the plant species (Lu et al. 2021). When the concentration of the contaminant is <500 ppm, an increase in growth parameters has been reported (Liu et al. 2018; Frachia et al. 2022), but a decrease in height and biomass in plants developed with 50-100 ppm Pb was also observed (Chaturvedi et al. 2020).



Figure 1. Weight of fresh weed biomass in soil contaminated and not contaminated by lead.

#### Plant growth promotion by rhizospheric bacteria

In the rhizosphere of three lead-tolerant weeds, 26 colonies of Gram-negative (76.9%) and Gram-positive (23.1%) bacteria were isolated on nutrient agar with 50 mg L<sup>-1</sup> of Pb, among which 92.3% grew in nutrient broth with 100 and 200 mg L<sup>-1</sup> of Pb, 76.9% in broth with 400 mg L<sup>-1</sup> of Pb, and 50% in broth with 800 mg L<sup>-1</sup> of Pb. In the rhizospheric soil of *E. colona* (L.) Link, 10 cultures of bacteria were obtained, and 60% of Gram-negative bacteria grew with 800 mg L<sup>-1</sup> of Pb. In the rhizospheric soil of Pb. In the rhizospheric soil of *C. corymbosus* Rottb., seven cultures of bacteria were obtained that grew with 400 mg L<sup>-1</sup> of Pb, and 71.4% with 800 mg L<sup>-1</sup> of Pb. In the rhizospheric soil of *S. halepense*, nine cultures of bacteria were obtained, among which 44.4% grew in broth with 400 mg L<sup>-1</sup> of Pb, and 22.2% in broth with 800 mg L<sup>-1</sup> of Pb.

Similar results were reported by Gorelova et al. (2022) and Manzoor et al. (2019), who isolated microorganisms in the rhizosphere of plants in soil contaminated with lead artificially and by anthropic activity, respectively. Manzoor et al. (2019) determined 62% of lead tolerant Gram-negative bacteria. Moreover, Liu et al. (2018) determined significant changes in the microbial population developed in a lead-contaminated site, an increase in the relative abundance of the bacterial genera *Flavisolibacter, Kaistobacter,* and *Pseudomonas,* as well as a decrease in the genera *Bacillus, Adhaeribacter, Pantibacter,* and *Paenibacillus.* The rhizosphere is the soil close to the roots (5 mm) that is directly influenced by root exudates, and where the largest population of microorganisms that interact with plants is found (Gamalero and Glick 2022).

All the microorganisms from bacterial consortium one (BC1) solubilized tricalcium phosphate and produced AIA (Table 1), characteristics that show *in vitro* growth promotion in plants. Phosphate solubilizing bacteria produce organic acids that destabilize the precipitated mineral and release soluble phosphorus to plants (Gavrilescu 2022). The increase in phosphorus availability favors plant growth and metabolite processes such as photosynthesis, energy

transfer, and carbohydrate utilization (Corrales et al. 2017). Likewise, rhizospheric microorganisms produce auxins such as IAA that induce the proliferation of the root system, cytokinins related to cell growth and regulation of hormones such as ethylene through the activity of the enzyme amino 1-carboxylate (ACC) deaminase, in addition to nitrogen fixation, phosphate solubilization, siderophore synthesis and ammonium production (Gavrilescu 2022).

Bacterial consortia	Bacterium code	Isolation source (Weeds)	Soluble phosphorus * (ppm)	AIA* (ppm)
BC1	1	<i>E. colona</i> (L.) Link	11.93±0.25	69.00±0.31
	3	<i>E. colona</i> (L.) Link	9.09±0.27	51.00±0.29
	4	<i>E. colona</i> (L.) Link	7.97±0.32	64.75±0.38
	6	<i>E. colona</i> (L.) Link	8.18±0.36	63.50±0.35
	7	<i>E. colona</i> (L.) Link	10.55±0.39	65.75±0.41
	9	<i>E. colona</i> (L.) Link	7.41±0.24	66.25±0.38
	17	C. corymbosus Rottb.	0	37.75±0.28
	11	C. corymbosus Rottb.	7.45±0.41	41.75±0.37
DOO	13	C. corymbosus Rottb.	7.40±0.34	54.25±0.28
BC2	12	C. corymbosus Rottb.	0	42.75±0.27
	14	C. corymbosus Rottb.	0	45.5±0.41
	15	C. corymbosus Rottb.	0	52.5±0.52
	20	S. halepense	0	0
BC3	25	S. halepense	6.88±0.39	40.25±0.38
	24	S. halepense	0	36.75±0.36
	26	S. halepense	7.24±0.33	36.10±0.39
	20	S. halepense	0	0
	24	S. halepense	0	0

Table 1. Solubilized phosphorus and indole acetic acid produced by lead-tolerant bacterial consortia.

\* Average of three repetitions.

# Influence of rhizospheric bacterial consortia on weed growth and phytotoxicity of phytoremediated soil

The consortia of rhizospheric bacteria positively influenced the growth of lead-tolerant weeds, observing increases of 14.32% (*E. colona* (L.) Link + BC1), 4.60% (*Cyperus corymbosus* Rottb. + BC2) and 4.14% (*Sorghum halepense* + BC3). Tukey's mean comparison test showed that the highest value of fresh biomass corresponded to the *E. colonum* plus BC1, with significant differences compared to the other treatments (Figure 2), a similar result to the reported by Manzoor et al. (2019) in the ornamental plants *Pelargonium hortorum* and *Mesembryanthemun*  *criniflorum* grown in soil artificially contaminated with 500-2,000 mg kg<sup>-1</sup> of Pb. The researchers showed that *Bacillus tequilensis* increased the biomass of both plants in soils with and without lead concentrations. There was an increase of 65.52% in *M. criniflorum* and 64.04% in *P. hortorum* in the soil with 2,000 mg kg<sup>-1</sup> of Pb.

Liu et al. (2018) showed that 100 ppm of lead did not affect the growth of *T. repens*; however, 500 ppm of lead significantly decreased root length (19.08%), plant height (25.92%), aerial biomass (25.1%) and root biomass (24.3%).



Figure 2. Fresh biomass of weeds at 90 days after inoculation of lead-tolerant bacterial consortia (BC). Different lowercase letters indicate significant differences between treatments (*P*<0.05).

The level of phytotoxicity was from severe to moderate in the soil where the lead-tolerant weeds were grown for 90 days, and it was from moderate to low in the soil with the lead-tolerant weeds inoculated with the bacterial consortia. The soil where the highest germination rate of radish seeds was reached (89.56%), and therefore the lowest level of phytotoxicity (low) was that corresponding to *E. colona* (L.) Link with BC1 (Table 2).

Table 2. Soil phytotoxicity level in Raphanus sativus L. seeds 90 days after sowing weeds with and without lead-tolerant bacteria consortia.

Treatments	Relative percentage of germination (RPG)	Average root elongation (cm)	Relative radicle growth (RRG)	Germination rate GI (%)	Phytotoxicity level
E. colona (L.) Link	79.17	3.22	60.64	48.01	Moderate
C. corymbosus Rottb.	77.08	3.1	58.38	44.99	Severe
Sorghum halepense	76.04	2.97	55.93	42.53	Severe
E. colona (L.) Link + BC1	92.7	5.13	96.61	89.56	Low
C. corymbosus Rottb. + BC2	88.54	4.63	87.19	77.2	Moderate
S. halepense + BC3	86.45	4.8	90.39	78.14	Moderate

# Parameters of lead phytoremediation in soil with the lower phytotoxicity

Lead phytoremediation parameters in soil with lower phytotoxicity, planted with *E. colona* (L.) Link increased with the inoculation of bacterial consortia. In the *E. colona* (L.) Link inoculated with BC1, the bioaccumulation factor

(BAC) was 0.1651 in the foliage, and 1.0250 in the roots; the translocation factor (TF) was 0.1611 compared to 0.0115, 1.0116 and 0.0114 in plants not inoculated with bacteria. In this context, the percentage of lead removal in the soil was 78.83% with *E. colona* (L.) Link inoculated with BC1, and 57.58% with uninoculated *E. colona* (L.) Link (Table 3).

Table 3. Parameters of phytoremediation of soil contaminated with lead by *Echinochloa colona* (L.) Link with and without bacterial consortium 1 (BC1).

Parameters	Values	
Initial lead into the ground (mg kg <sup>-1</sup> )	800.0	
Lead in foliage of plants without bacteria (mg kg <sup>-1</sup> )	9.23	
Lead in plant roots without bacteria (mg kg <sup>-1</sup> )	808.5	
Lead in soil from plants without bacteria (mg kg <sup>-1</sup> )	339.38	
Lead in foliage of plants with bacteria (mg kg-1)	132.13	
Lead in plant roots with bacteria (mg kg-1)	820.0	
Lead in soil from plants with bacteria (mg kg <sup>-1</sup> )	169.38	
BAC of foliage of plants without bacteria	0.0115	
BAC of roots without bacteria	1.0106	
TF of foliage and roots without bacteria	0.0114	
Lead removal in plants without bacteria (%)	57.58	
BAC of plant foliage with bacteria	0.1651	
BAC of roots with bacteria	1.0250	
TF of foliage and roots with bacteria	0.1611	
Lead removal in plants with bacteria (%)	78.83	

TF<1=Plant that does not translocate metal and is a phytostabilizer. BAC>1 and TF<1=Phytostabilizer plant, and 1  $\leq$  BAC<10=Bioaccumulator plant.

The concentration of lead accumulated in the roots was higher than the foliage of the E. colona (L.) Link plants with and without bacteria, a superiority that coincides with the studies of Gorelova et al. (2022) and Lu et al. (2021). In this regard, Gorelova et al. (2022) guantified 1.57-7.38 mg kg<sup>-1</sup> of Pb in roots, and 0.14–0.88 mg kg<sup>-1</sup> of Pb in foliage of Echinochloa frumentacea, and concluded that this species is not suitable for remediation, because it accumulates and removes only a small amount of lead (0.18–0.94 g ha<sup>-1</sup>), compared to other plants such as Sorghum halepense with 107–378 g ha<sup>-1</sup>. After lead enters the soil, it remains mostly in divalent forms (oxides and sulfides). Precipitated forms of lead phosphate [Pb<sub>3</sub>(PO<sub>4</sub>)<sub>2</sub>] and lead carbonate (PbCO<sub>2</sub>) are found in the root, which are difficult to transport to the aerial part (Lu et al. 2021). Plants respond to lead stress in different ways, but higher BAC is common in roots compared to foliage (Lu et al. 2021). These researchers determined a BAC range of 0.931-1.062 in the root of Plantago asiatica L. and 0.074-0.087 in the leaves, a result that demonstrated that the root is the main organ for absorption and accumulation of lead.

The bioaccumulation factor (BAC) of lead in the roots of *Echinochloa colona* (L.) Link with and without bacteria was greater than one and the translocation factor (TF) was less than one, a result that demonstrated this plant species is a phytostabilizer of lead, due to plants with a BF>1 and TF<1 are appropriate for phytostabilization (Rigoletto et al. 2020). The bioaccumulation or bioconcentration factor estimates the absorption capacity of metals in plants, and the translocation factor evaluates the transport of metal from the root to the foliage (Rigoletto et al. 2020). The TF values of 0.1611 and 0.0114 in E. colona (L.) Link with and without bacteria were lower than the range of 0.061–0.819 reported by Lu et al. (2021) in Taraxacum mongolicum and Plantago asiatica L. developed with 5% of lead. According to Rigoletto et al. (2020), plants with a translocation factor less than one are not capable of transporting the contaminant from the roots to the foliage and are used for the phytostabilization of the contaminant. In this way, they immobilize lead and avoid the contamination of groundwater.

The lead content decreased in the phytoremediated soil with and without bacteria, a decrease previously reported by Lu et al. (2021) and Gorelova et al. (2022). At the end of the evaluation period of this study, 57.58% removal of the contaminant was reached in the soil of the plants where

bacteria were not applied, and 78.83% when bacteria were applied. The range of percent lead removal is greater than the 13.01–31.89% reported by Lu et al. (2021) in soils contaminated with 2–5% lead and remediated with *Artemisia capillaris* and *Plantago asiatica* L.

Microorganisms favor the phytoremediation of contaminated soil directly by reducing the toxic effect and increasing the metal absorbed in the roots and foliage (Manzoor et al. 2019), and indirectly through the promotion of plant growth (Gavrilescu 2022). Growth-promoting microorganisms increase the solubility of lead because they decrease the pH (Manzoor et al. 2019), adsorb lead on their surface and precipitate (Zhu et al. 2022), capture lead by the biosorption mechanism, modify the bioavailability of lead through the production of siderophores, phosphate solubilization, production of organic acids and exopolysaccharides (Shah and Daverey 2020), and transform lead into inert forms such as lead sulfate (PbSO<sub>4</sub>) and lead sulfides (PbS) (Rigoletto et al. 2020).

Due to the formation of insoluble precipitates, the availability of lead is very low in the soil. In this context, citric and malic

acid are used to improve the accessibility of the heavy metal in the soil and the translocation from the root to the foliage (Rigoletto et al. 2020). Therefore, phosphatesolubilizing bacteria of the genera *Pseudomonas* and *Acinetobacter* favored the availability of lead, as well as greater absorption by the root system of lead-tolerant plants (Gupta et al. 2018).

# Identification of the genus of bacteria in the consortium with the greatest influence on phytoremediation

The genera *Pseudomonas* and *Acinetobacter* were identified among the bacteria in consortium one (Table 4). In this regard, it coincides with the microorganisms recovered by Manzoor et al. (2019) (*Pseudomonas*) and Zhu et al. (2022) (*Acinetobacter*). The bacteria reported as tolerant to lead, with potential for phytoremediation of soil contaminated with heavy metal, belong to the genera *Klebsiella*, *Pseudomonas*, *Sporosarcina*, *Microbacterium*, *Staphylococcus*, *Bacillus* (Manzoor et al. 2019), *Flavisolibacter*, *Kaistobacter* (Liu et al. 2018), as well as the species *Acinetobacter calcoaceticus* (Zhu et al. 2022), *Delftia acidovorans* and *Azonexus caeni* (Abdelrahman and Younggy 2022).

 Table 4. Characteristics of the Pseudomonas and Acinetobacter genera identified in bacterial consortium one (BC1) isolated from lead-tolerant E. colona (L.) Link.

Biochemical tests	Pseudomonas*	Acinetobacter**
Catalase test	+	+
Oxidase test	+	
Motility test	+	
Acid production (TSI)	K/K	A/K
Nitrate reduction	-	
Urea hydrolysis	-	
Starch hydrolysis	-	+
Indole production	-	
Citrate use	+	+
Lysine decarboxylation (LIA)	K/A	K/A
Gelatin hydrolysis	-	
Aesculin hydrolysis	-	-

\*Bacteria code: 1, 3, 7, 9; \*\*Bacteria code: 4, 6. K: Alkalinity, A: Acidity, -: Negative, and +: Positive.

#### CONCLUSION

The weeds *Echinochloa colona* (L.) Link, *Cyperus corymbosus* Rottb. and *Sorghum halepense* with associated rhizobacteria were efficient in the remediation of soil

contaminated with lead. 78.83% of heavy metal removal was achieved with *E. colona* (L.) Link plus *Pseudomonas* spp. and *Acinetobacter* spp. (BC1), compared to 57.80% with *E. colona* (L.) Link without bacteria. *Pseudomonas* spp.

and Acinetobacter spp. were favored by the root exudates of lead-tolerant plants, which are used as nutrients. This association between weeds and plant growth-promoting bacteria can be applied as an alternative to achieve environmental sustainability through the remediation of soil impacted by several heavy metals.

#### REFERENCES

Abdelrahman A and Younggy K (2022) Isolation of Pb(II)-reducing bacteria and demonstration of biological Pb(II) reduction to metallic Pb. Journal of Hazardous Materials 423: 126975. https://doi.org/10.1016/j. jhazmat.2021.126975

Acosta L and Bustamante D (2020) Caracterización de microorganismos oxidantes del azufre y su potencial para la recuperación de suelo sódico con la aplicación de azufre (Tesis para optar el título profesional). Universidad Nacional Pedro Ruiz Gallo. Lambayeque, Perú. 82 p.

Álvarez-López C, Osorio-Vega W, Diéz-Gómez M et al (2014) Biochemical characterization of rhizosphere microorganisms from vanilla plants with potential as biofertilizers. Agronomía Mesoamericana 25: 225-241. https://doi.org/10.15517/am.v25i2.15426

Amadi N and Gbosidom V (2022) Assessment of heavy metal uptake ability of *Echinochloa colona* in orange peel amended soil. International Journal of Advanced Research in Biological Sciences 9: 40-48. https://ijarbs.com/pdfcopy/2022/jan2022/ijarbs5.pdf

Chaturvedi R, Favas P, Pratas J et al (2020) Harnessing *Pisum* sativum-Glomus mosseae symbiosis for phytoremediation of soil contaminated with lead, cadmium, and arsenic. International Journal of Phytoremediation 23: 279-290. http://doi.org/10.1080/15226514. 2020.1812507

Contreras H and Carreño C (2018) Eficiencia de la biodegradación de hidrocarburos de petróleo por hongos filamentosos aislados de suelo contaminado. Revista de Investigación Científica UNTRM: Ciencias Naturales e Ingeniería 1: 27-33. https://doi.org/10.25127/ucni.v1i1.269

Contreras-Pinto L, Valencia C, De la Fuente N et al (2016) Estudio de absorción, acumulación y potencial para la remediación de suelo contaminado por plomo usando *Ambrosia ambrosioides*. Investigación y Desarrollo en Ciencia y Tecnología de Alimentos 1: 244-250. http: //www.fcb.uanl.mx/IDCyTA/files/volume1/1/2/41.pdf

Corrales L, Caycedo L, Gómez M et al (2017) *Bacillus* spp: una alternativa para la promoción vegetal por dos caminos enzimáticos. NOVA 15: 45-65. https://revistas.unicolmayor.edu.co/index.php/nova/article/view/588

Covarrubias S and Peña J (2017) Contaminación ambiental por metales pesados en México: Problemática y estrategias de fitoremediación. Revista Internacional de Contaminación Ambiental 33: 7-21. https://doi.org/10.20937/RICA.2017.33.esp01.01

Dipak P and Sinha S (2017) Isolation and characterization of phosphate solubilizing bacterium *Pseudomonas aeruginosa* KUPSB12 with antibacterial potential from river Ganga, India. Annals of Agrarian Science 15: 130-136. http://doi.org/10.1016/j.aasci.2016.10.001

Frachia C, Navarro V, da Silva W et al (2022) *Inga uruguensis* response to lead: effects on growth and nitrogenous compounds. Rodriguésia 73: e01652020. https://doi.org/10.1590/2175-7860202273063

Gamalero E and Glick B (2022) Recent advances in bacterial

amelioration of plant drought and salt stress. Biology 11: 437. https:// doi.org/10.3390/biology11030437

Gavrilescu M (2022) Enhancing phytoremediation of soils polluted with heavy metals. Current Opinion in Biotechnology 74: 21-31. https://doi.org/10.1016/j.copbio.2021.10.024

Gorelova S, Muratova A, Zinicovscaia I et al (2022) Prospects for the use of *Echinochloa frumentacea* for phytoremediation of soils with multielement anomalies. Soil Systems 6: 27. https://doi.org/10.3390/ soilsystems6010027

Gupta P, Kumar V, Usmani Z et al (2018) Phosphate solubilization and chromium (VI) remediation potential of *Klebsiella* sp. strain CPSB4 isolated from the chromium contaminated agricultural soil. Chemosphere 192: 318-327. https://doi.org/10.1016/j.chemosphere.2017.10.164

Lakhal D, Boutaleb N, Bahlaouan B et al (2017) Mixture experimental design in the development of a bio fertilizer from fish waste, molasses and scum. International Journal of Engineering Research & Technology 6: 588–594. https://www.ijert.org/mixture-experimental-design-in-the-development-of-a-bio-fertilizer-from-fish-waste-molasses-and-scum

Liu A, Wang W, Zheng X et al (2022) Improvement of the Cd and Zn phytoremediation efficiency of rice (*Oryza sativa*) through the inoculation of a metal-resistant PGPR strain. Chemosphere 302: 134900. https://doi.org/10.1016/j.chemosphere.2022.134900

Liu C, Lin H, Dong Y et al (2018) Investigation on microbial community in remediation of lead-contaminated soil by *Trifolium repens* L. Ecotoxicology and Environmental Safety 165: 52-60. https://doi. org/10.1016/j.ecoenv.2018.08.054

Lu N, Li G, Sun Y et al (2021) Phytoremediation potential of four native plants in soils contaminated with lead in a mining area. Land 10: 1129. https://doi.org/10.3390/land10111129

Manzoor M, Gul I, Ahmed I et al (2019) Metal tolerant bacteria enhanced phytoextraction of lead by two accumulator ornamental species. Chemosphere 227: 561-569. https://doi.org/10.1016/j. chemosphere.2019.04.093

MINAM – Ministerio del Ambiente (2017) Aprueban Estándares de Calidad Ambiental (ECA) para Suelo. Decreto Supremo N° 011-2017-MINAM. Peru. https://www.gob.pe/institucion/minam/normaslegales/3693-011-2017-minam

Nakbanpote W, Meesungnoen O and Prasad M (2016) Potential of ornamental plants for phytoremediation of heavy metals and income generation. Bioremediation and Bioeconomy 2016: 179-217. https://doi.org/10.1016/B978-0-12-802830-8.00009-5

Rigoletto M, Calza P, Gaggero E et al (2020) Bioremediation methods for the recovery of lead-contaminated soils: A Review. Applied Sciences 10: 3528. https://doi.org/10.3390/app10103528

Rodríguez R, Cuéllar L, Maldonado C et al (2016) Efectos nocivos del plomo para la salud del hombre. Revista Cubana de Investigaciones Biomédicas 35: 251-271. https://www.medigraphic.com/cgi-bin/new/ resumen.cgi?IDARTICULO=70505

Sabreena, Hassan S, Bhat S et al (2022) Phytoremediation of heavy metals: an indispensable contrivance in green remediation technology. Plants 11: 1255. https://doi.org/10.3390/plants11091255

Salas-Marcial C, Garduño-Ayala M, Mendiola-Ortiz P et al (2019) Fuentes de contaminación por plomo en alimentos, efectos en la salud y estrategias de prevención. Revista Iberoamericana de Tecnología Postcosecha 20: 1-15. https://www.redalyc.org/articulo. oa?id=81359562002 Salazar E and Nieves B (2005) *Acinetobacter* spp. Aspectos microbiológicos, clínicos y epidemiológicos. Revista de la Sociedad Venezolana de Microbiología 25: 64-71. http://caelum.ucv.ve/ojs/index. php/rev\_vm/article/view/444

Shah V and Daverey A (2020) Phytoremediation: a multidisciplinary approach to clean up heavy metal contaminated soil. Environmental Technology & Innovation 18: 100774. https://doi.org/10.1016/j. eti.2020.100774

Silva J, Avello C, Matamoro F et al (1999) Antimicrobial resistance of different *Acinetobacter baumanni* biotypes isolated in the northern region of Chile. Revista médica de Chile 127: 926-934. https://doi.org/10.4067/S0034-98871999000800006

Zhu X, Li X, Shen B et al (2022) Bioremediation of lead-contaminated soil by inorganic phosphate-solubilizing bacteria immobilized on biochar. Ecotoxicology and Environmental Safety 237: 113524. https://doi. org/10.1016/j.ecoenv.2022.113524

